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Hongyun Hu, Dunkui Chen, Huan Liu, Yuhan Yang, Hexun Cai, Junhao Shen, Hong Yao

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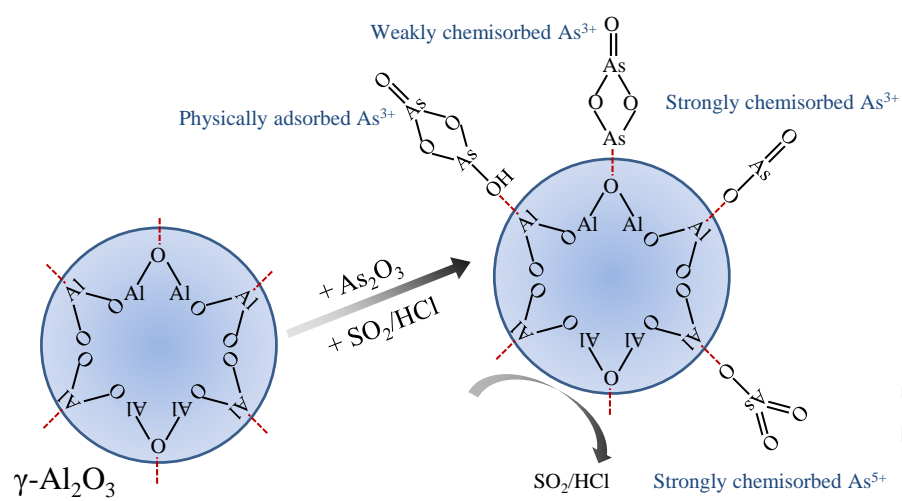
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Graphical abstract



1 **Adsorption and reaction mechanism of arsenic vapors over γ -Al₂O₃ in the**
2 **simulated flue gas containing acid gases**

3 Hongyun Hu^a, Dunkui Chen^a, Huan Liu^{a,b}, Yuhan Yang^a, Hexun Cai^a, Junhao Shen^a,
4 Hong Yao^{a,b,*}

5 ^a *State Key Laboratory of Coal Combustion, School of Energy and Power Engineering,*
6 *Huazhong University of Science and Technology, Wuhan 430074, China*

7 ^b *Department of New Energy Science and Engineering, School of Energy and Power*
8 *Engineering, Huazhong University of Science and Technology, Wuhan 430074, China*

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*Corresponding author. Tel & Fax: +86-27-87545526(O)

E-mail: hyao@mail.hust.edu.cn (Prof. Hong Yao)

10 **Abstract**

11 Arsenic emission from fuel combustion and metal smelting flue gas causes
12 serious pollution. Addition of sorbents is a promising way for the arsenic capture from
13 high temperature flue gas. However, it is difficult to remove arsenic from
14 SO₂/HCl-rich flue gas due to the competitive reaction of the sorbents with arsenic and
15 these acid gases. To solve this problem, arsenic adsorption over γ -Al₂O₃ was studied
16 in this work to evaluate its adsorption mechanism, resistance to acid gases as well as
17 regeneration behavior. The results show that γ -Al₂O₃ had good resistance to acid gases
18 and the arsenic adsorption by γ -Al₂O₃ could be effectively carried out at a wide
19 temperature range between 573 and 1023 K. Nevertheless, adsorption at
20 higher-temperature (like 1173 K) led to the decrease of surface area and the
21 rearrangement of crystal structure of γ -Al₂O₃, reducing the active sites for arsenic
22 adsorption. The adsorption of arsenic was confirmed to occur at different active sites
23 in γ -Al₂O₃ by forming various adsorbed species. Increasing temperature facilitated
24 arsenic transformation into more stable chemisorbed As³⁺ and As⁵⁺ which were
25 difficult to remove through thermal treatment regeneration. Fortunately, the
26 regeneration of spent γ -Al₂O₃ could be well performed using NaOH solution.

27

28 **Keywords:** Arsenic adsorption; γ -Al₂O₃; Acid gases; Species; Regeneration

29 1. Introduction

30 Arsenic is extremely toxic and the release of arsenic has raised great
31 environmental concerns (Sarkar and Paul, 2016). Nowadays, positive measures have
32 been taken in China to strictly control the arsenic emission, especially for the major
33 arsenic-emitting industries, like fuel combustion and metal smelting (He et al., 2013).
34 Due to the high volatilization, arsenic is predominantly evaporated into the high
35 temperature flue gas in the form of vapor-phases (such as As_2O_3) (Bashaa et al., 2008;
36 Shen et al., 2015). Arsenic vapors tend to be enriched in fine particles as temperature
37 decreased in the cooling process of the flue gas, which is difficult to captured by
38 normally used air pollution control devices (Hu et al., 2016; Liu et al., 2016). On the
39 other hand, the presence of arsenic in the flue gas has negative effect for the recovery
40 of SO_2 in the metal smelting (Dalewski, 1999). Moreover, it may also cause the
41 deactivation of the catalysts of the widely used selective catalytic reduction system for
42 NO_x control (Hu et al., 2016). Therefore, it's essential for the capture of arsenic
43 vapors from the high temperature flue gas.

44 Previous literatures, including our research, have demonstrated that interactions
45 between arsenic vapors and Al, Ca and/or Fe-compounds enhanced arsenic
46 stabilization in coarse particles, increasing arsenic trapping efficiency (Hu et al., 2015;
47 Zhang et al., 2016). Thus, it provides a promising way for arsenic capture from the
48 flue gas through the addition of Al, Ca and/or Fe-based sorbents. Usually, arsenic is
49 coexisted with sulfur in most of coals and metal minerals (Shen et al., 2015; Lattanzi
50 et al., 2008). As a result, arsenic-rich flue gas probably contains a high concentration

51 of SO₂. Among the arsenic capture sorbents, CaO was widely investigated and
52 showed good capacity for the adsorption of arsenic vapors (Li et al., 2007; Tian et al.,
53 2016). However, CaO had low resistance to SO₂, and arsenic sorption was strongly
54 suppressed by the reaction of SO₂ with CaO (Seams and Wendt, 2007; Chen et al.,
55 2015). So far, it is still a challenge to find more effective and suitable sorbents for the
56 adsorption of arsenic at the presence of SO₂.

57 Comparing with CaO, active alumina has large surface area and proper pore
58 structures. Using of active alumina for the arsenic sorption from liquids has been
59 successfully carried out (Han et al., 2013; Önnbya et al., 2014). Several studies have
60 also confirmed that arsenic could be captured from the high temperature flue gas by
61 alumina (Contreras et al., 2009; Luo et al., 2011). However, the adsorption of arsenic
62 under the influence of acid gases had been scarcely investigated. Comparing with
63 CaO, alumina contained stronger surface acid sites and weaker basic sites (Zhao et al.,
64 2011). Based on this, alumina might show better resistance to acid gases than CaO
65 although the adsorption of SO₂ by alumina at low temperatures was observed (Xie et
66 al., 2014). In addition to the adsorption atmosphere, the sorption behavior was
67 generally determined by the reaction temperature (Ninomiya et al., 2009; Wang et al.,
68 2013). The change of the operating temperature might alter the pathway for the
69 adsorption of SO₂ as well as arsenic vapors. In view of these points, the present study
70 observe the arsenic adsorption by alumina under the influence of acid gases at various
71 temperatures, and the mechanism regarding the arsenic adsorption and reaction over
72 alumina was clarified.

73 In particular, arsenic adsorption mechanism on active alumina ($\gamma\text{-Al}_2\text{O}_3$) was
74 studied at temperatures ranging from 573 K to 1323 K. The experiments were
75 conducted in the simulated flue gas containing arsenic vapors within/without acid
76 gases (SO_2/HCl). The differences in arsenic sorption capacity were discussed
77 considering the effects of adsorption temperature and the presence of acid gases in the
78 flue gas. To further understand the mechanism regarding the arsenic adsorption, the
79 species of the arsenic captured on the alumina were determined. For the recycling of
80 the $\gamma\text{-Al}_2\text{O}_3$, regeneration of the spent alumina was carried out through both thermal
81 treatment and NaOH solution leaching process.

82 2. Experimental procedures

83 2.1 Materials

84 Synthetic $\gamma\text{-Al}_2\text{O}_3$ used in this work was purchased from Aladdin (shanghai,
85 China). Before experiments, the alumina particles were carefully sieved and the size
86 fraction between 200 and 300 μm was selected. To provide constant arsenic vapors,
87 arsenic was firstly transformed into $\text{AsH}_3(\text{g})$ by using a hydride generator and then
88 was oxidized to form arsenic vapors at 973 K (as shown in Fig. SM-1 in
89 Supplementary Material (SM)). In the hydride generator, $\text{AsH}_3(\text{g})$ was produced by
90 mixing of NaAsO_2 solution (dissolved in 1 v/v% HCl solution), HCl solution and
91 KBH_4 solution. The concentration of $\text{AsH}_3(\text{g})$ was constantly maintained through the
92 precise control of the NaAsO_2 solution flow rate using a peristaltic pump. Thereafter,
93 the produced $\text{AsH}_3(\text{g})$ was carried by pure argon to mix with N_2 , O_2 and acid gases
94 (SO_2 and HCl) before entering the reactor. All the reagents used here are of high

95 purity grade.

96 **2.2 Adsorption Experiments**

97 Fig. SM-1 illustrates the schematic diagram of arsenic adsorption system. The
98 system mainly consisted of two parts, including a mixing reactor for $\text{AsH}_3(\text{g})$
99 production and a two-stage reactor for $\text{AsH}_3(\text{g})$ oxidation and arsenic vapors
100 adsorption. Firstly, the simulated flue gas with $\text{AsH}_3(\text{g})$ was prepared as stated before.
101 The flue gas was heated at 973 K in the first stage of the two-stage reactor. After the
102 oxidation of $\text{AsH}_3(\text{g})$, the simulated flue gas was measured, comprising 86 v/v% N_2 , 6
103 v/v% O_2 and 8 v/v% $\text{H}_2\text{O}(\text{g})$ and 200ppm $\text{As}_2\text{O}_3(\text{g})$. In the case of observing the
104 effects of acid gases on the arsenic adsorption, the SO_2 or HCl was input into the
105 reactor and the concentration was kept at 2000ppm and 500ppm, respectively.

106 Secondly, the simulated flue gas was sent to the second stage of the two-stage
107 reactor and the adsorption of arsenic vapors was conducted at temperatures ranging
108 from 573 to 1323 K. Before arsenic adsorption, a certain amount of $\gamma\text{-Al}_2\text{O}_3$ was
109 placed in the sorbent holder and was preheated in the flue gas without $\text{As}_2\text{O}_3(\text{g})$. Then,
110 the hydride generator was operated and the arsenic vapors were present in the
111 simulated flue gas. The reactions between $\gamma\text{-Al}_2\text{O}_3$ and the simulated flue gas
112 compositions were taken place in the sorbent holder and the details of the sorbent
113 holder were explained elsewhere (Chen et al., 2015).

114 Finally, the reacted alumina after adsorption was cooled to the room temperature
115 in the arsenic free flue gas, after which the products were collected for the subsequent
116 analysis. The observation of arsenic sorption capacity for high-temperature treated

117 alumina as well as regenerated alumina was carried out in the same way as for the raw
118 γ -Al₂O₃.

119 **2.3 Regeneration of spent alumina**

120 The regeneration of spent alumina was tried by using two methods in this study.
121 On one hand, the adsorbed arsenic was removed from the spent alumina through
122 thermal treatment in arsenic-free atmosphere. The spent alumina products obtained at
123 low temperatures from 573 to 1023 K were tested, which were sent to heat at 1173 K
124 for 10 min in N₂ or air atmosphere.

125 On the other hand, the regeneration of the spent alumina was carried out by
126 immersing the used sorbents in diluted NaOH solution at 298 and 348 K, respectively.
127 The regenerated alumina was washed by deionized water and dried at 378 K. The
128 performance of the NaOH-treated alumina for arsenic capture was tested at 573 K for
129 7 cycles.

130 **2.4 Analytical methods**

131 The adsorbed arsenic in the reacted alumina was leached out by 20 v/v% HCl
132 solution and determined by atomic fluorescence spectroscopy (AFS). Besides, the
133 speciation of the adsorbed arsenic was measured by using AFS coupled with high
134 performance liquid chromatography (HPLC). Before the speciation detection, arsenic
135 should be also transformed into solution by using 0.5 mol/L phosphoric acid as
136 leaching solution. The leached various forms of arsenic (arsenite (As³⁺) and arsenate
137 (As⁵⁺)) were separated in the analytical column (Hamilton PRP-X100) and the
138 concentration of each species of arsenic was determined by AFS. Each experiment

139 was repeated for several times to make sure the results believable. To compare the
140 arsenic adsorption capacity in the flue gas with and without acid gases, a difference
141 ratio (β) was defined and calculated as Ep. (1).

$$\beta = \frac{m_2 - m_1}{m_1} \times 100\% \quad (1)$$

142 where, m_1 means arsenic capture capacity in the flue gas without SO_2/HCl , mg/g
143 sorbent, m_2 means arsenic capture capacity in the flue gas with SO_2/HCl , mg/g
144 sorbent.

145 The properties of the activated alumina before and after arsenic adsorption were
146 comparatively observed. X-ray powder diffraction (XRD, X'Pert PRO, PANalytical
147 B.V.) was applied to analyze the crystalline structure change of the alumina. The
148 measurement of BET specific surface area was conducted via N_2 isothermal
149 adsorption (Micromeritics ASPS 2020).

150 3. Results and Discussion

151 3.1 Effects of acid gases on the arsenic adsorption by $\gamma\text{-Al}_2\text{O}_3$

152 Fig. 1 presents the arsenic adsorption by $\gamma\text{-Al}_2\text{O}_3$ in the flue gas with and without
153 SO_2 at a retention time of 5 min. The results showed that $\gamma\text{-Al}_2\text{O}_3$ showed good
154 resistance to SO_2 at each temperature from 573 to 1323 K. Compared with CaO (Chen
155 et al., 2015), the presence of SO_2 in the simulated flue gas had little impact on the
156 arsenic capture by $\gamma\text{-Al}_2\text{O}_3$. It might be attributed to the large surface area of $\gamma\text{-Al}_2\text{O}_3$.
157 $\gamma\text{-Al}_2\text{O}_3$ could offer enough active sites for the adsorption for both As_2O_3 and SO_2 in
158 the short reaction time.

159 To further confirm the suspicion, $\gamma\text{-Al}_2\text{O}_3$ was exposed in the arsenic-containing

160 flue gas for longer retention time. The difference ratios (as defined in section 2.2)
161 regarding the arsenic adsorption at the absence or presence of SO_2 were displayed in
162 Fig. 2. The presence of SO_2 generally had a negative effect on the arsenic adsorption,
163 which became remarkable with the retention time increasing from 5 to 90 min. It was
164 probable that the adsorption of both As_2O_3 and SO_2 was taken place in the same active
165 site of $\gamma\text{-Al}_2\text{O}_3$. The active sites were gradually occupied in the long-time adsorption
166 process. The lack of active sites led to the competitive adsorption of As_2O_3 and
167 SO_2 over $\gamma\text{-Al}_2\text{O}_3$. As a result, the adsorption capacity of arsenic was reduced. Even
168 so, $\gamma\text{-Al}_2\text{O}_3$ had much better resistance to SO_2 than CaO .

169 Similarly, the effect of HCl on the arsenic adsorption over $\gamma\text{-Al}_2\text{O}_3$ was observed.
170 The experiments were performed in 5 min and the results (shown in Fig. SM-2)
171 compared the effect of HCl on CaO and $\gamma\text{-Al}_2\text{O}_3$ for the adsorption of arsenic. From
172 the results, arsenic capture by CaO was strongly affected by HCl while the capture of
173 arsenic by $\gamma\text{-Al}_2\text{O}_3$ was slightly affected. Moreover, the effect of HCl showed
174 different tendencies at various temperatures for both CaO and $\gamma\text{-Al}_2\text{O}_3$. The
175 adsorption of HCl by CaO was mainly through the chemisorption (Tongamp et al.,
176 2009) which was enhanced with the temperature increasing from 573 K to 1023 K.
177 Therefore, the adsorption of arsenic by CaO was greatly suppressed at 1023 K.
178 However, the presence of HCl had little impact or even positive effect on the arsenic
179 adsorption by CaO at higher temperatures. It was widely observed that CaO particles
180 were easily sintered at high temperatures (Liu et al., 2010; Hu et al., 2013). The
181 surface characteristics of the sintering CaO particles might be improved under the

182 influence of HCl, increasing the capture capacity of CaO (Wu et al., 2014). In contrast,
183 the adsorption of HCl on the surface of $\gamma\text{-Al}_2\text{O}_3$ was mainly through the physical
184 adsorption (Vigue et al., 1998), which was effective at 573 K. The acidity of HCl was
185 stronger than that of SO_2 and the adsorption of HCl on the surface of $\gamma\text{-Al}_2\text{O}_3$ is not
186 conducive to the capture of arsenic. Nevertheless, the physical sorption of HCl was
187 inhibited at high temperatures and the effect of HCl became progressively smaller
188 with the increase of the adsorption temperatures.

189 On the other hand, an interesting phenomena was found in Fig. 1. Unlike CaO,
190 the capacity of $\gamma\text{-Al}_2\text{O}_3$ for the arsenic adsorption was almost the same at
191 temperatures from 573 to 1023 K. It was attractive especially for the effective arsenic
192 capture at low temperatures. However, arsenic adsorption was partly suppressed at
193 1173 and 1323 K and the similar phenomena was also found for the arsenic capture by
194 CaO (Chen et al., 2015). Compared with CaO, arsenic capture by $\gamma\text{-Al}_2\text{O}_3$ was less
195 affected at high temperatures and the mechanism of $\text{As}_2\text{O}_3(\text{g})$ adsorption over $\gamma\text{-Al}_2\text{O}_3$
196 should be further addressed.

197 **3.2 Mechanism of $\text{As}_2\text{O}_3(\text{g})$ adsorption over $\gamma\text{-Al}_2\text{O}_3$**

198 Fig. 3 depicts the arsenic adsorption capacity over $\gamma\text{-Al}_2\text{O}_3$ at various retention
199 time as a function of temperature. From the results, the increase of retention time had
200 few effect on the arsenic adsorption at temperatures between 573 K and 1023 K.
201 However, arsenic adsorption at higher temperatures was strongly affected by the
202 retention time. The adsorption capacity was remarkably decreased at 1173 and 1323 K
203 with the increasing of the retention time from 1 min to 10 min. The inhabitation of

204 arsenic adsorption at high temperature was further confirmed by determining arsenic
205 adsorption over thermally treated Al_2O_3 (shown in Table SM-1). The adsorption
206 capacity was sharply decreased after heating Al_2O_3 for 10 min and longer time. In
207 contrast, the adsorption capacity was hardly affected when $\gamma\text{-Al}_2\text{O}_3$ was heated at 573
208 K and 873 K even for 60 min (not shown here).

209 Generally, $\gamma\text{-Al}_2\text{O}_3$ is one kind of the metastable “transition” alumina structural
210 polymorphs which is unstable during high temperature treatment (Busca, 2014). As
211 shown in Table SM-2, the specific surface area of $\gamma\text{-Al}_2\text{O}_3$ particles was reduced in the
212 high temperature adsorption process. In detail, few changes in the specific surface
213 area were found as Al_2O_3 heated at 1023 K. However, large decrease in specific
214 surface area was observed for samples as the adsorption temperature was raised to
215 1173 and 1323 K. On the other hand, the chemical phase of the Al_2O_3 was changed at
216 these high temperatures. Fig. SM-3 shows the XRD patterns of the alumina heated at
217 1273 K as a function of heating time. The presence of new diffraction peaks (such as
218 at 2θ values of 31.8° and 32.9°) revealed that Al_2O_3 was structurally rearranged in the
219 high temperature adsorption process. Specifically, $\gamma\text{-Al}_2\text{O}_3$ was transformed to
220 $\theta\text{-Al}_2\text{O}_3$ in the heating process as supported by previous studies (Kwak et al., 2011).
221 Based on preliminary analysis, the inhabitation of arsenic adsorption at high
222 temperatures was attributed to the change of the physical characteristics as well as the
223 crystal structure of Al_2O_3 . The decrease of specific surface area might lead to few
224 active sites available for the adsorption of arsenic. And the structural rearrangements
225 of $\gamma\text{-Al}_2\text{O}_3$ was supposed to change the adsorption sites for arsenic.

226 To clarify the pathway for arsenic adsorption over γ -Al₂O₃, detailed studies were
227 carried out to identify the speciation of the captured arsenic (from the flue gas with
228 and without SO₂). The distribution of As³⁺ fraction in the reaction products were listed
229 in Table 1 and the rest of arsenic was in the form of As⁵⁺. Arsenic was predominantly
230 in the form of As³⁺ in the reaction products obtained at temperatures from 573 K to
231 873 K. Increasing of the adsorption temperature favored arsenic capture in the form of
232 arsenates (As⁵⁺). The presence of SO₂ in the flue gas slightly affected the pathway for
233 arsenic adsorption over Al₂O₃, which tended to facilitate arsenic distribution in the
234 form of As³⁺.

235 Like the adsorption of SO₂ (Zhao et al., 2012), arsenic adsorption could take
236 place at different active sites over γ -Al₂O₃ by forming various adsorbed species. Fig.
237 SM-4 gives several possible model of adsorbed arsenic species. At low temperature
238 like 573 K, the interactions of γ -Al₂O₃ and moisture resulted in the formation of
239 hydroxyl groups (\bullet OH). Then arsenic vapors could be adsorbed on these groups by
240 forming physically adsorbed species (Fig. SM-4(a)). Meanwhile, arsenic could be
241 captured to form weakly chemisorbed species (Fig. SM-4(b)). Increasing temperature
242 enhanced the strongly chemisorption of arsenic both in the form of As³⁺ and As⁵⁺ (Fig.
243 SM-4(c) and (d)), which probably had good thermal stability. These speculations were
244 consistent with the results regarding the detection of the adsorbed arsenic's stabilities
245 which were discussed in section 3.3.

246 3.3 Regeneration of spent γ -Al₂O₃

247 As stated before, a large fraction of arsenic might be adsorbed through physical

248 adsorption at low temperatures in the form of As^{3+} . The removal of physically
249 adsorbed species from sorbents was usually proposed by using thermal treatment
250 method (Yang et al., 2014; Hamzehlouyan et al., 2016). Based on this, the removal of
251 arsenic from the reacted Al_2O_3 was conducted at 1173 K in arsenic-free atmosphere
252 (air and N_2 , respectively). Both of the removal efficiency of arsenic and the speciation
253 of the remained arsenic was observed. As shown in Fig. 4, the removal efficiency
254 differed in the products obtained at various adsorption temperatures. More arsenic
255 was released from the reacted Al_2O_3 obtained at low temperatures like 573 and 723 K.
256 Arsenic adsorbed at high temperatures was difficult to release. Compared with arsenic
257 removal in N_2 , more arsenic was stabilized in the Al_2O_3 by forming As^{5+} under air
258 atmosphere.

259 The discrepancy in thermal stability of the captured arsenic further confirmed the
260 presence of various species of the adsorbed arsenic over Al_2O_3 . The high fraction of
261 released arsenic for products obtained at 573 K and 723 K suggested that arsenic
262 adsorption was mainly through physically adsorption and weakly chemisorption at
263 low temperature. The As^{3+} in the products obtained at high temperatures was assigned
264 to the strongly chemisorbed As^{3+} which could be well stabilized even heated at 1173
265 K. The strongly chemisorbed As^{5+} generally stable at the experimental temperature as
266 evident from the large fraction of remained arsenic in the form of As^{5+} .

267 Considering the poor removal efficiency of captured arsenic during thermal
268 treatment, the regeneration of spent $\gamma\text{-Al}_2\text{O}_3$ was tried through the leaching process of
269 NaOH solution. After regeneration, the $\gamma\text{-Al}_2\text{O}_3$ was exposed in the simulated flue gas

270 for the adsorption of arsenic at 573 K for several recycles. The results in Fig. 5
271 indicated that the regeneration of γ -Al₂O₃ was effectively conducted by using alkaline
272 solution leaching. In contrast, operation at 348 K showed a better performance than
273 that at 298 K in the subsequent cycling experiments.

274 The regeneration temperature could affect the extraction of arsenic from the
275 reacted γ -Al₂O₃. As shown in Table 2, regeneration at 348 K could maintain a high
276 extraction fraction of arsenic. Less arsenic was extracted at 298 K and the extraction
277 fraction was further decreased with the increasing of regeneration cycles. As a result,
278 more arsenic was remained in the regenerated γ -Al₂O₃ and less active sites were left
279 for the adsorption of arsenic. The regeneration at 348 K enhanced the interaction
280 between NaOH solution and γ -Al₂O₃, promoting the arsenic extraction. The similar
281 phenomena was found for the regeneration of spent magnetic nanomaterials for
282 arsenic removal from liquids (Zhang et al., 2010). However, the regeneration by using
283 NaOH solution probably increased the surface basic sites of γ -Al₂O₃, which might
284 reduce the resistance to acid gases. It's essential to comprehensively evaluate the
285 properties of the regenerated γ -Al₂O₃ by using alkaline solution.

286 4. Conclusions

287 The mechanism were illuminated regarding the arsenic adsorption over γ -Al₂O₃
288 in the simulated flue gas containing acid gases. γ -Al₂O₃ showed good performance for
289 the adsorption of arsenic at a wide temperature range from 573 to 1023 K which was
290 hardly affected at the presence of acid gases in a certain retention time. However, in
291 the long-time adsorption process, acid gases competed with arsenic vapors to react

292 with γ -Al₂O₃, suppressing the adsorption of arsenic. The high temperature adsorption
293 was also not conducive for the arsenic adsorption due to the damage of the pore
294 structure as well as the rearrangement of the crystal structure of γ -Al₂O₃.

295 The adsorption of arsenic over γ -Al₂O₃ was a complicated process with arsenic
296 adsorbed at various active sites in γ -Al₂O₃. The formation of different arsenic species
297 was confirmed which showed various thermal stability. Although the arsenic capacity
298 was almost the same as adsorption conducted at temperatures from 573 to 1023 K, the
299 increasing of adsorption temperature enhanced the arsenic transformation into more
300 stable chemisorbed As³⁺ and As⁵⁺. Moreover, the spent γ -Al₂O₃ was successfully
301 regenerated through the leaching process by NaOH solution. The regenerated γ -Al₂O₃
302 at 348 K showed good performance for the arsenic adsorption for recycling for 7
303 times.

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Tables list

Table 1 Distribution of As^{3+} fraction (%) in the reaction products obtained from the simulated flue gas with and without SO_2

Table 2 Extraction fraction (%) of arsenic from the reaction products after multiple cycles of adsorption

Table 1 Distribution of As³⁺ fraction (%) in the reaction products obtained from the simulated flue gas with and without SO₂

Flue gas characteristic	573 K	723 K	873 K	1023 K	1173 K	1323 K
Without SO ₂	100	87.21	83.62	47.48	19.4	0
Within SO ₂	93.0	94.47	91.47	58.26	19.5	3.65

Table 2 Extraction fraction (%) of arsenic from the reaction products after multiple cycles of adsorption

Extracting temperature	Number of cycle					
	1	2	3	4	5	6
298 K	63.4	45.7	35.3	28.1	25.4	22.3
348 K	87.6	90.1	85.2	83.9	85.3	81.1

Figure captions

Fig. 1. Arsenic adsorption capacity by alumina in the flue gas with/without SO₂ at temperatures from 573 to 1323 K

Fig. 2. The difference ratio of arsenic capture in the simulated flue gas with/without SO₂ at various retention time

Fig. 3. Arsenic adsorption capacity by alumina at temperatures from 573 to 1323 K

Fig. 4. Arsenic distribution in the reaction products after thermal treatment at 1173 K (Left column for treatment in air and right column for treatment in N₂)

Fig.5. Multiple cycles of arsenic adsorption by alumina regenerated at various temperatures

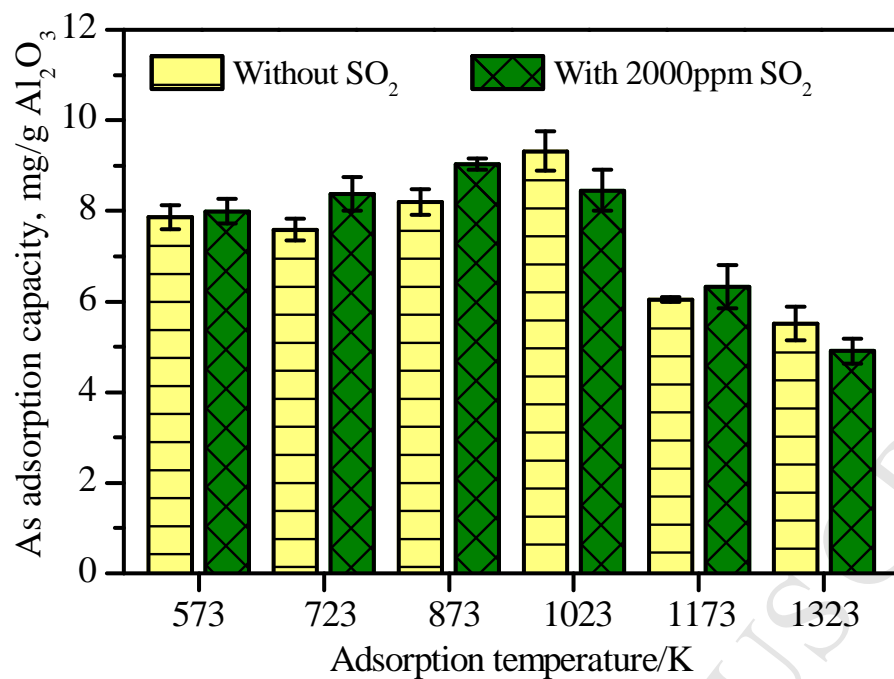


Fig. 1. Arsenic adsorption capacity by alumina in the flue gas with/without SO₂ at temperatures from 573 to 1323 K

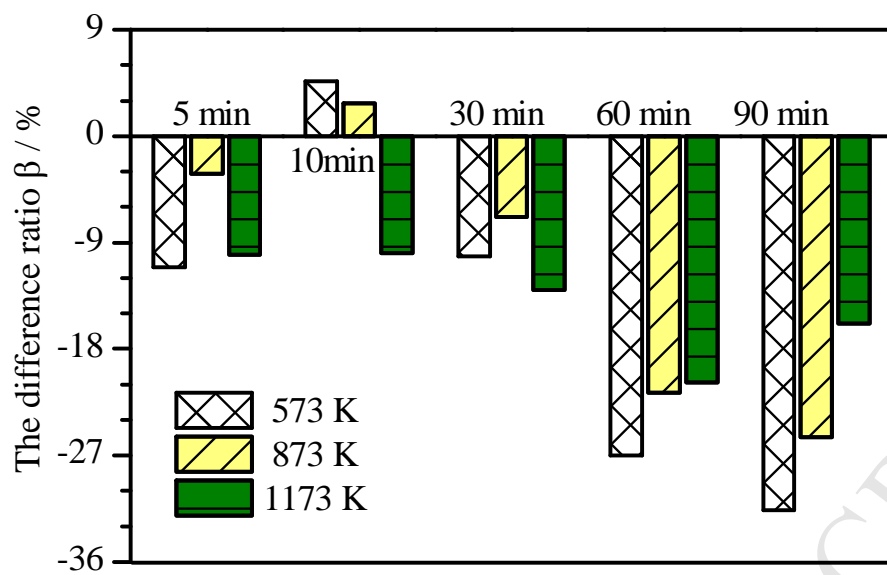


Fig. 2. The difference ratio of arsenic capture in the simulated flue gas with/without SO_2 at various retention time

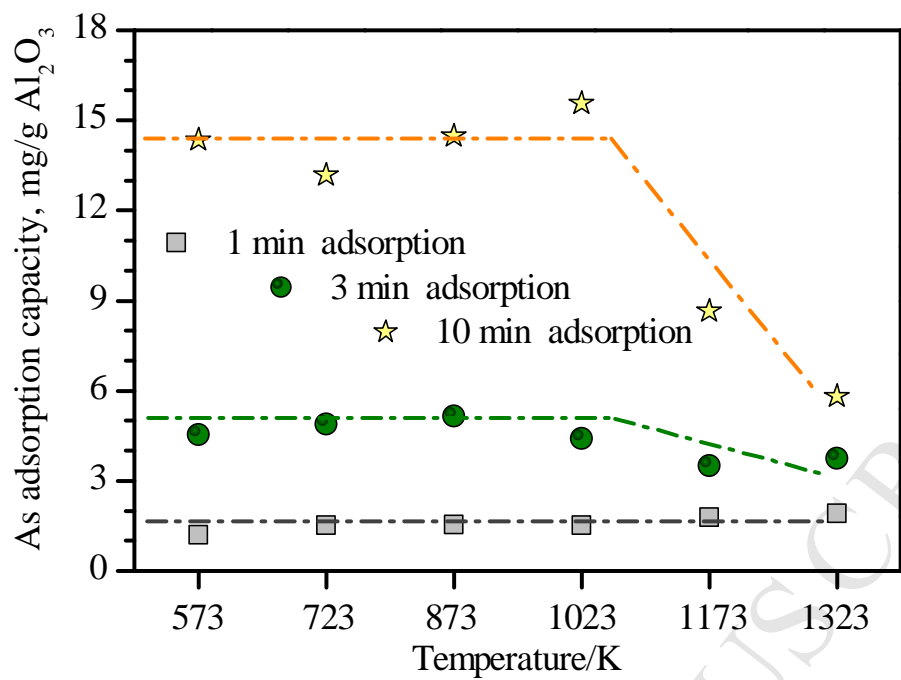


Fig. 3. Arsenic adsorption capacity by alumina at temperatures from 573 to 1323 K

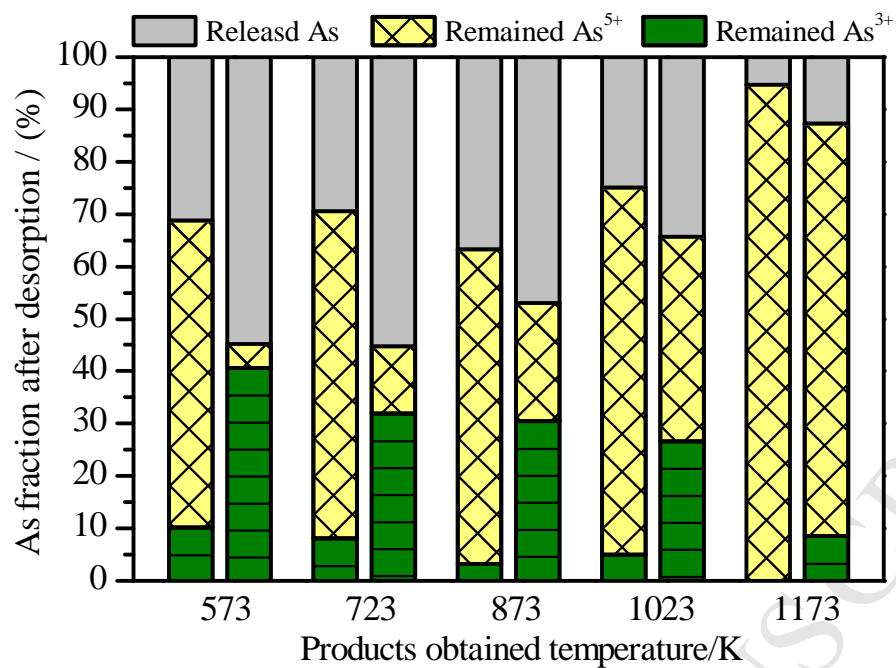


Fig. 4. Arsenic distribution in the reaction products after thermal treatment at 1173 K

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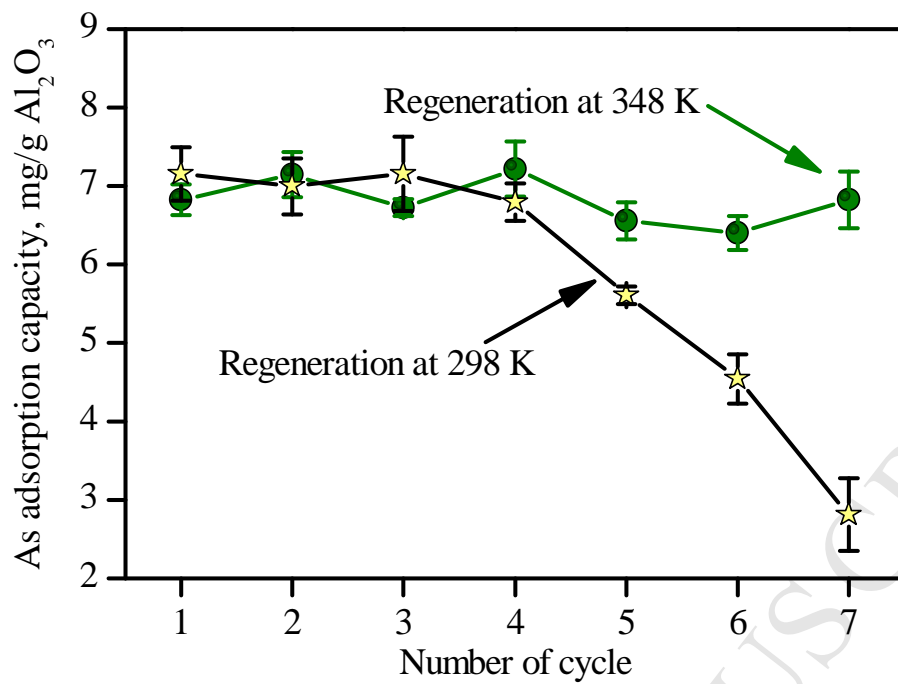


Fig.5. Multiple cycles of arsenic adsorption by alumina regenerated at various temperatures

Highlights

- As capture by γ -Al₂O₃ was effective at a wide temperature range from 573 to 1023 K.
- γ -Al₂O₃ showed good resistance to acid gases in the process for arsenic adsorption.
- As was adsorbed at different active sites in γ -Al₂O₃ by forming various species.
- As capture was inhibited by the change of γ -Al₂O₃ structure at high temperatures.
- The regeneration of spent γ -Al₂O₃ could be well performed using NaOH solution.